Carbon burial and storage in tropical salt marshes under the influence of sea level rise

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HIGHLIGHTS
• 210Pb-dated sediment cores from tropical salt marshes showed marine transgression.
• Corg concentration, stock and burial varied widely within and among study sites.
• Corg stocks in tropical saltmarshes are as high as in other blue carbon ecosystems.
• Marine transgression caused lower Corg sediment concentrations and stocks.
• Sea level rise effects on Corg burial rates are masked by other global change impacts.

Abstract
Coastal vegetated habitats can be important sinks of organic carbon (Corg) and mitigate global warming by sequestering significant quantities of atmospheric CO2 and storing sedimentary Corg for long periods, although their Corg burial and storage capacity may be affected by on-going sea level rise and human intervention. Geochemical data from published 210Pb-dated sediment cores, collected from low-energy microtidal coastal wetlands in El Salvador (Jiquilisco Bay) and in Mexico (Salada Lagoon; Estero de Urias Lagoon; Sian Ka’an Biosphere Reserve) were revisited to assess temporal changes (within the last 100 years) of Corg concentrations, storage and burial rates in tropical salt marshes under the influence of sea level rise and contrasting anthropization degree. Grain size distribution was used to identify hydrodynamic changes, and δ13C to distinguish terrigenous sediments from those accumulated under the influence of marine transgression. Although the accretion rate ranges in all sediment records were comparable, Corg concentrations (0.2–30%), stocks (30–465 Mg ha⁻¹, by extrapolation to 1 m depth), and burial rates (3–378 g m⁻² year⁻¹) varied widely within and
1. Introduction

Coastal ecosystems such as tidal wetlands and seagrasses (collectively known as coastal blue carbon ecosystems) are recognized for their capacity to remove CO₂ from the atmosphere and incorporate it, along with trapped plant materials, into their sediments (Howard et al., 2014). Tidal wetlands are important in climate change research because of their sensitivity to sea-level rise and their ability to sequester CO₂ from the atmosphere, which might otherwise contribute to global warming. The capacity of long-term organic carbon burial by wetland sediments primarily depends on the ratio between inputs (organic matter produced in situ and ex situ) and outputs (decomposition and erosion; Scholz, 2010). Oxygen penetrates and is depleted by aerobic respiration within the upper few millimeters of soil, whereas anaerobic metabolism, especially sulfate reduction, dominates at greater depths (Alongi and Brinkman, 2011). Since this process is less efficient than surface aerobic metabolism, organic carbon is stored almost indefinitely until exposed by shoreline erosion or soil desiccation.

Organic carbon (Corg) storage in coastal wetlands can be significantly disturbed by anthropogenic interventions and/or by climate-induced factors, such as increasing air and water temperatures (associated with atmospheric CO₂ increase), change in precipitation patterns, and sea-level rise (Smook et al., 2013). Sea level is perhaps the most urgent concern for intertidal ecosystems as, to persist, their surface elevations must increase in keeping with rising sea level (Chmura et al., 2003). Otherwise, the Corg sink capacity of these ecosystems may be compromised, since the buried organic matter may become exposed to conditions favorable to decomposition and remineralization to gaseous form (Breithaupt et al., 2012).

Tropical salt marshes are usually found along semi-sheltered low-energy coastlines, located behind the mangrove fringe, in the highest possible topographic position within the tidal range, so they are intermittently flooded by medium to high tides, at least once fortnightly. High evaporation and relatively infrequent flooding favor the formation of hypersaline soils that can be colonized by halophyte grasswet vegetation (e.g. Batis maritima, Salicornia pacifica) that can tolerate inundation with sea water and high soil salinity (Costa et al., 2009; Flores-Verdugo et al., 2007). The ecosystem services associated to salt marshes includes fish production, carbon storage and coastal protection; however, they have been historically perceived as wastelands, and have thus experienced intense human impacts (e.g. land reclamation for agriculture, urban development, salt production, waste disposal, livestock grazing, and shrimp aquaculture) for which 25% to 50% of their global historical coverage has been lost (Mcowen et al., 2017 and references therein). These ecosystems are so under-valued that there are no statistical data on their extent or loss rate due to anthropogenic transformation in Mexico or El Salvador.

Since 2013, the United Nations Framework Convention on Climate Change established the need to promote sustainable management and conservation of greenhouse gases sinks and reservoirs, including coastal ecosystems. Nowadays, there is an increasing interest to better understand the role of natural coastal wetlands (particularly mangroves, seagrasses and tidal marshes) as long-term carbon sinks, and their role in climate-change mitigation (Howard et al., 2017). However, few quantitative data on Corg accumulation and preservation in tropical tidal marshes are available. In addition, despite the efforts to predict the effects of sea level rise on coastal carbon sinks by using numerical models, short-term Corg degradation experiments, or salinity gradients as analogue for marine transgression (e.g. Kirwan and Mudd, 2012; Van de Broek et al.; 2016; Whittle and Gallego-Sala, 2016) this issue is still not clear, mostly owing to the scarcity of field studies that show temporal changes in both ongoing sea level rise and Corg accumulation.

With the goal to contribute to the knowledge of the long-term organic carbon burial capacity and carbon storage of tropical salt marshes, and to the discussion on how sea level rise might affect these ecosystem services, published geochemical data from 210Pb-dated sediment cores, collected from four low-energy microtidal tropical salt marsh areas (three from Mexico and one from El Salvador) where marine transgression has been confirmed, were revisited to assess the temporal changes Corg concentrations (%), Corg stocks (Mg ha⁻¹, stored within 1 m depth and within a centennial scale), and Corg burial rates (g m⁻² year⁻¹). This work provides quantitative information on the long-term Corg burial and storage in tropical salt marshes, derived from the retrospective evaluation of Corg accumulation in four different geographical areas under the influence of ongoing sea level rise (as shown by geochemical proxies for marine transgression, such as C:N ratios and isotopic organic carbon composition). Our field observations highlight the contribution of tropical salt marshes to CO₂ sequestration and sustain the importance of their preservation to mitigate global warming.
the sea) is surrounded by a mangrove forest colonized by R. mangle, L. racemosa and A. germinans. The climate is warm and humid with average temperature of 25 °C (7.5 °C to 39 °C; SMN, 2014) and summer rains (annual precipitation 700–1300 mm; INEGI, 2009). The coastal plain is composed of Quaternary alluvial soils (predominantly regosol) overlying Tertiary igneous rocks, e.g. andesite, rhyolite, dacite, trachyte, rhyodacite, basalt and tuff (Alba-Cornejo et al., 1979).

The Salada Lagoon is a small (51 ha) and shallow (≤ 3 m depth) coastal lagoon in the state of Veracruz (Gulf of Mexico, Fig. 1), near the Laguna Verde Nuclear Power Plant. The lagoon does not have a permanent freshwater supply (except rainfall runoff), and until the mid-1970s its connection with the sea was limited, which caused desiccation owing to high evaporation rates and, consequently, high salinities (up to 70; Rodríguez Castañeda, 1994). Nowadays, the lagoon receives the discharge of cooling waters from the power plant through a channel built in 1981 (Morales-Lozano, 2011). The channel established a permanent connection between the lagoon and the sea, decreasing salinities to values comparable to seawater (37; Rodríguez Castañeda, 1994). The climate is warm-humid with summer rains and an average annual temperature that ranges from 21 °C to 28 °C (CONABIO, 2012). The mangrove species present are R. mangle, L. racemosa and A. germinans. The area is characterized by the presence of igneous extrusive rocks, e.g. andesite, basalt, basic tuff and breccia; with phaeozem and vertisol as the main soil types.

Sian Ka’an is a large (~528,000 ha) and complex hydrological system on the Mexican Caribbean coast, registered as a World Heritage Site, UNESCO Biosphere Reserve, Wildlife Protection Area and RAMSAR site. It lies on a partially emerged coastal karstic plain facing a 120 km-long barrier reef, and includes two large shallow bays surrounded by mangroves (R. mangle, A. germinans and L. racemosa), numerous sinkholes (“cenotes”) and a vast extension of tropical deciduous forests. Much of the Reserve is limestone of Pleistocene origin. The climate is tropical with summer rains (mean annual rainfall 1300 mm) and occasional cyclones, with a mean annual temperature of 26.5 °C (extreme values ranging from 4.5 °C to 40.5 °C).

3. Methods

Most sediment records referred here have already been used to describe environmental processes related to global change (Amaya-Monterrosa et al., 2014; Ruiz-Fernández et al., 2016; Carnero-Bravo et al., 2016; Bojórquez-Sánchez et al., 2017), where further methodological details can be found.

3.1. Sampling

Sediment push cores were collected with PVC tubes (10 cm diameter, 50 cm length) between May 2012 and April 2013 (Fig. 1). The cores
were cut into 1 cm thick sections, and samples were freeze-dried and ground to powder in a porcelain mortar (except those for grain size analysis). The sediment cores were collected in remote areas, far away from human and cattle access, in relatively flat sites with presence of glassworts (*S. pacifica* and/or *B. maritima*), between 10 and 50 m behind the landward edge of mangrove areas. The shorter distance between the closest tidal creek and the sampling sites in Jiquilisco Bay (El Salvador) were 560 m for core ELS I and 270 m for core ELS II; and in Mexico, 100 m for EU I and 150 m for EUIII (Estero de Urías Lagoon); 320 m for SA I and 150 m for SA II (Salada Lagoon); and 50 m for ISCUL and 10 m for PUPA (Sian Ka’an Biosphere Reserve).

3.2. Laboratory analysis

Grain size distribution (sand, silt and clay fractions, %) was determined by laser diffraction with a Malvern Mastersizer 2000. The elemental concentrations and isotopic composition of organic C and N were determined on carbonate-free sediment samples, using elemental
analysers coupled to isotope ratio mass spectrometers. At Stanford University Stable Isotope Laboratory (cores from Estero de Urrías Lagoon and Jiquilisco Bay) sediments were placed into silver capsules and decarbonated with 6% H2SO4 whereas at CICIMAR (cores from Salada Lagoon) and GEOTOP-UQAM (cores from Sian Ka'an Biosphere Reserve) sediments were treated with 1 M HCl and freeze dried before analysis. Isotopic results are presented in standard delta notation (δ), with δ13C reported relative to the Vienna Pee Dee Belemnite (VPDB) carbon standard and δ15N relative to air. Total 210Pb (210Pb) activities were determined through its radioactive descendant 210Po by alpha spectrometry (Ortec Ametek 576A spectrometer), by the method described by Ruiz-Fernández and Hillaire-Marcel (2009), whereas supported 210Pb (210Pbsup) activities were determined through the measurement of 228Ra by low background gamma-spectrometry in well-type HPGe detectors (ORTEC). Excess 210Pb (210Pbex) was calculated as the difference between 210Pb and 210Pbsup. The 210Pb-derived chronologies were calculated with the constant flux (CF) model (Robbins, 1978) which assumes a constant flux of 210Pb to the sediments, and variable accumulation rates (the assumptions and formulae for dating, mass accumulation rates (MAR) and sediment accretion rates (SAR), are described in detail in Sanchez-Cabeza and Ruiz-Fernández, 2012). Age, MAR and SAR uncertainties were estimated by Monte Carlo simulation with 30,000 iterations, by using the 210Pb dating Excel workbook described in Sanchez-Cabeza et al. (2014). The 210Pb-derived dating data is available as Supplementary information.

Analytical quality control included accuracy assessment through the analysis of reference materials, and precision assessment by replicate analysis of a single sample (n = 6). Results were within the reported range of the certified values. The variation coefficients were <5% for grain size distribution, <3% for C and N isotopic composition, and <10% for 210Pb activities.

3.3. Data analysis

In order to evaluate the influence of sea level rise on Corg accumulation, a correlation analysis between Corg concentrations and δ13C (as a proxy for marine transgression) was performed. All data was used as a single data set; normality of the data distribution was tested by the Kolmogorov-Smirnov test (p < 0.01) and the statistical significance of the correlation analysis was determined by using a Student’s t-test (p < 0.05).

4. Results

4.1. 210Pb activity profiles and 210Pb-derived mass accumulation rates

The 210Pb profiles for all the analyzed cores are collated in Fig. 2 from previously published work (Amaya-Monterrosa et al., 2014; Ruiz-Fernández et al., 2016; Carnero-Bravo et al., 2016; Bojórquez-Sánchez et al., 2017). All sediment records spanned at least 100 years. The ranges of both mass accumulation rate (MAR, 0.03 ± 0.01 to 0.88 ± 0.65 cm−2 year−1) and sediment accretion rates (SAR, 0.04 ± 0.01 to 1.03 ± 0.77 cm year−1) were comparable among the sediment cores (Table 1).

SAR ranges were compatible with those reported for mangrove areas in (a) the Gulf of Mexico (0.24 cm year−1 in Terminos Lagoon, Lynch et al., 1989; 0.2–0.4 cm year−1 in the lagoons of Terminos, Chelem and Celestun, in the Western Yucatan Peninsula, Gonnetea et al., 2004); (b) the Marismas Nacionales System in the Pacific coast of Mexico (0.08 ± 0.01 to 0.36 ± 0.02 cm year−1; Ruiz-Fernández et al., 2018); (c) the global estimate of 0.25–0.36 cm year−1 by Twilley et al. (1992); and (d) the general mangrove forest worldwide range of 0.5–0.7 cm year−1 by Alongi (2012). The estimated recent rates of sea level rise (mm year−1, derived from sediment accretion rates, under the assumption that these ecosystems accrete at a similar rate to sea level rise) are comparable across the four study sites, ranging from 3.4 ± 0.5 to 3.7 ± 0.3 in Jiquilisco (period 2010–2013, Amaya-Monterrosa et al., 2014); 2.8 ± 0.1 to 3.9 ± 0.1 in Estero de Urrías Lagoon (1990–2012, Ruiz-Fernández et al., 2016), 3.0 ± 0.7 to 4.5 ± 0.6 in Sian Ka’an BR (2010–2013, Carnero-Bravo et al., 2016), and 1.7 ± 0.2 to 4.3 ± 0.3 in Laguna Salada (2007–2013, Bojórquez-Sánchez et al., 2017). These results are comparable to the mean sea level rise rate recorded at the nearest tide gauge of each study site, for the available periods. They are also compatible with the mean sea level rise trends estimated from tide gauges in Mexico and El Salvador (CO-OPS, 2013; no uncertainties provided) and the global mean SLR of 3.0 ± 0.7 mm year−1 from the analysis of sea-level data of a global network of tide gauges (from 1993 to 2010, Hay et al., 2015). Despite that sea level rise rates derived from SAR reflect local (relative) sea level rise trends, our estimations are also compatible with the global mean sea level rise (GMSL) of 3.4 ± 0.4 mm year−1 derived from multi-mission satellite altimetry data (1993–2014; Abdall et al., 2017 and references therein). It should be noticed that altimetry-derived GMSL indicates the general state of sea level across the oceans and not at any specific location, and that the accuracy of
the altimetry-derived GMSL values is affected by the possibility of altimeter drifts, changes of satellite missions, and the geophysical corrections needed, for which satellite altimetry is less sensitive to long term (decadal) variability than tide-gauge measurements (Nerem et al., 2018). Nonetheless, tide gauge data in our study sites is very limited, and for this reason SAR-derived sea level rise trends represents a valuable alternative to reconstruct long term SLR trends.

4.2. Sediment characterization

Sediments in tropical salt marshes are a mixture of upland runoff, wetland organic production and suspended marine particles that are transported by tides, and settle and accumulate around the plant stems or following microtopography. Thus, sediment grain size distribution might help to identify hydrodynamic changes, and δ13C variations can help to trace the terrestrial or marine character of the sediments.

Fig. 3. Characterization of sediment cores from Jiquilisco Bay (ELS) in El Salvador, and from Estero de Urias Lagoon (EU), Salada Lagoon (SA) and Sian Ka’an BR (ISCUL, PUPA) in Mexico. Data from Amaya-Monterrosa et al. (2014), Ruiz-Fernández et al. (2016), Carnero-Bravo et al. (2016), and Bojórquez-Sánchez et al. (2017) except δ13C and C:N from ELS and SA.
4.2.1. Grain size distribution

All cores were composed mainly of sand and silt, with small percentages of clay (<25%). The depth profile of the grain size distribution (Fig. 3) was almost uniform in cores ELS I, EU I, EU III and PUPA; in cores SA I and SA II it showed a cyclic variation between silt and sand content; and in core ELS II it showed a transition from sandy to silty sediments. In Sian Ka’an BR, the deeper segments of both sediment cores contained mangrove remnants (roots and wood chunks) accounting for ~5% of the section mass below 20 cm depth in PUPA, and ~30% below ~10 cm in IScul. In fact, mineral material was almost absent from the section of IScul enriched by plant debris, so grain size analysis was not feasible.

4.3. Organic carbon

4.3.1. Organic carbon concentrations and trends

C_{org} concentrations ranged from <1% to 30% (Fig. 3). Highest C_{org} concentrations were usually observed in cores with higher contents of
fine sediment, in agreement with the common observation that fine-grained particles have a larger surface area by unit volume and adsorb Corg more efficiently than coarser particles (Loring and Rantala, 1992). However, highest Corg concentrations were observed in the section of core ISCUL with high plant debris content (for which there is no information regarding grain size).

4.3.2. Organic carbon provenance

Sources of organic matter are commonly inferred from sediment bulk properties such as elemental composition, C:N ratios and stable isotopes of C and N. Terrestrial organic matter, composed mainly of vascular plant debris, is assumed to have C:N > 20 (Meyers, 1997), $\delta^{13}C$ between $-25\%$ and $-28\%$ (Burdige, 2005) and $\delta^{15}N$ around $+0.4 \pm 0.9\%$ (Peterson and Howard, 1987). Marine-derived organic matter is characterized by C:N < 10 (Parsons, 1975), $\delta^{13}C$ from $-17\%$ to $-24\%$ (Gearing et al., 1984) and $\delta^{15}N$ between 3‰ and 12‰ (Wada and Hattori, 1990). Thus, organic matter in coastal sediments should present C:N, $\delta^{13}C$ and $\delta^{15}N$ values between the ranges for terrestrial and marine end members. Ranges in our cores (Table 1) clearly show the mixed nature (terrestrial and marine) of organic matter in these sediments.

In all cores, depth profiles showed decreasing C:N ratios and increasing $\delta^{13}C$ values towards the surface (Fig. 3), indicating an increasing supply of marine organic matter (Fig. 4), which support that these low-lying tropical salt marshes are already reflecting marine transgression due to recent sea level rise (Ruiz-Fernández et al., 2016; Carnero-Bravo et al., 2016).
4.3.3. Organic carbon burial and storage

As none of the depth profiles for C<sub-org</sub> concentration showed the typical exponential decay caused by microbial degradation, it was impossible to evaluate the effects of diagenesis by using first-order mineralization models (e.g. Berner, 1964; Middelburg, 1989). The C<sub-org</sub> burial rate (i.e. deposition at the sediment surface) was estimated by multiplying C<sub-org</sub> concentrations by mass accumulation rates, assuming that C<sub-org</sub> concentrations represent those accumulated at the surface layer at the time of deposition, as if diagenesis effects were negligible.

C<sub-org</sub> burial rates ranged from 2.8 ± 0.9 to 378.0 ± 33.8 g m<sup>−2</sup> year<sup>−1</sup> (Table 1) and changed considerably within and among the study sites, with highest values in core ISCUL (Sian Ka’an BR), and lowest in the sandy sediments of ELS I (Jiquilisco Bay) and PUPA (Sian Ka’an BR). C<sub-org</sub> burial rates also changed widely with time, but time-dependent C<sub-org</sub> burial rate profiles for the two cores at each study site were comparable, excepting for cores PUPA and ISCUL, which showed opposite time-dependent profiles (Fig. 5D). With the exception of core ISCUL, C<sub-org</sub> burial rates in all cores gradually increased from the 1900s to the 1950s; subsequently, their patterns were diverse, with more abrupt and intermittent changes, particularly in cores from Estero de Urias (EU, Fig. 5B) and Salada (SA, Fig. 5C) lagoons, which are the most anthropized areas among the study sites.

C<sub-org</sub> stock (i.e. the amount of organic carbon stored per unit area down to a defined sediment depth) was calculated every 10 cm, from surface down to 40 cm depth, as the sum of the product of dry bulk density, C<sub-org</sub> concentrations and thickness of each core section; the uncertainties of C<sub-org</sub> stock calculations were estimated by quadratic uncertainty propagation (Fig. 6). The largest and smallest total C<sub-org</sub> stock values (surface-40 cm depth) were found in cores from a single site (Jiquilisco Bay) with a value ~16-fold larger in ELS II than in ELS I (Table 2). In both cores, the surface (0–10 cm) C<sub-org</sub> stocks were ~20% higher than in the underlying segment (Fig. 6), reflecting either a slight increase of C<sub-org</sub> concentrations in the uppermost layer or a small C<sub-org</sub> degradation with depth. In contrast, in cores EU I, SA I, PUPA and ISCUL, the surface C<sub-org</sub> stocks were between ~30% and ~60% lower than the segment immediately below, which might be explained by a lower recent C<sub-org</sub> supply. Except in ISCUL core, the C<sub-org</sub> stocks in the deepest segments (20–30 cm and 30–40 cm) were very similar, suggesting that below 20 cm C<sub-org</sub> degradation is negligible. However, in the ISCUL core (Sian Ka’an BR) the C<sub-org</sub> stock in the deepest segment was almost 40% higher than the overlying one.

Under the assumption of negligible C<sub-org</sub> degradation with depth, we used the average C<sub-org</sub> stock of the 20–40 cm depth segments to extrapolate the C<sub-org</sub> stock to 1 m depth, in order to compare our results with those from studies where the C<sub-org</sub> stock is calculated following the “blue carbon initiative” methodology (Howard et al., 2014). The C<sub-org</sub> stock values at 1 m in our sites ranged from 29.9 ± 0.4 to 464.8 ± 6.5 Mg ha<sup>−1</sup> (Table 2). In addition, taking into account that the time frame was different among the salt marshes owing to differences in mass accumulation rates, the C<sub-org</sub> stock was also calculated for the past 50 and 100 years (Table 2). At our sites, C<sub-org</sub> stocks accumulated during the past 50 years usually accounted for more than half of the stock accumulated during the past 100 years, except in ISCUL, where it accounted for only ~25%.

5. Discussion

Considering the diverse environmental settings and anthropization of these study sites, it is remarkable that the sedimentary records were so similar regarding grain size distribution, C:N ratios and δ<sup>13</sup>C trends, which implies the occurrence of similar sedimentary and geochemical processes at most sites.
5.1. Energy of the depositional environments

Grain size distribution varied notably along the records ELS II, SA I and SA II (Fig. 3), denoting hydrodynamic changes in the salt marsh environments. In core ELS II, the transition from sandy (below 25 cm depth) to silty sediments (surface to 25 cm depth) occurred >100 years ago, indicating a shift from a higher- to lower-energy environment, which was attributed to sea level rise, since recent sections contained remains of planktonic foraminifera carried by tidal currents (Amaya-Monterrosa et al., 2014). Rising sea level increases the probability and intensity of flooding low-lying coastal regions, and contributes to the redistribution of sediment along sandy coasts. The attenuation effect of the mangroves on the energy of the waves would promote that coarser particles are retained in the lower reaches of the intertidal zone, while the colonization of former upland areas by salt-resistant vegetation would promote that fine particles carried out by the tides are deposited in new flooded areas, where eventually would develop an onlap sequence, in which coarser sediments are covered by finer ones (Levin and King, 2017). In Salada Lagoon, the low sand content at 10 to 20 cm in core SA I and at 10 to 30 cm in core SA II could be related to periods of desiccation and inflilling. There is no information on the timeline and factors that might have promoted such events in the Salada Lagoon in the past, but a similar recent siltation process in a nearby lagoon is attributed to high evaporation caused by the seasonal closure of the lagoon mouth, acting together with the high erodibility of the surrounding terrain during the rainy season (Plata, 2005). The increase in sand content in the uppermost layers of both cores indicated a reversal from inflilling conditions to a higher-energy environment between the 1970s and 1980s, caused by the opening of the channel that conducts the cooling waters from the Laguna Verde nuclear power plant to the sea, which keeps the lagoon permanently connected to open sea waters.

In the ISICUL core from Sian Ka’an, the transition from sediments enriched with plant debris (10 to 50 cm depth) to silty grained sediments (surface to 10 cm depth) indicated that the current salt marsh area was formerly a mangrove forest. This transition was likely due to significant changes in environmental conditions (e.g. flooding, erosion, subsidence, soil hypersalinzation) that are generally followed by alterations in the vigor or zonation of mangrove trees, which may include widespread tree mortality (Jimenez et al., 1985). At the Sian Ka’an BR ISICUL station, a probable reason for this change may have been hypersalinzation caused by sea level rise.

5.2. \( C_{org} \) concentrations, burial rate and stock

The departure from exponential decay trends in most of the \( C_{org} \) depth profiles (Fig. 3B) could be indicative of non-steady \( C_{org} \) supply or non-steady degradation. However, taking into account the variations in sediment grain size distribution and in the provenance of the sediments (recent marine transgression) we considered that variations in \( C_{org} \) inputs are the most plausible explanation for these profiles. Non-steady-state conditions may be fostered by a combination of factors, such as tidal advection and temporal changes in supply rates and quality of organic sedimentation (Alongi et al., 1996). The almost constant \( C_{org} \) concentrations in some core segments (e.g. in ELS I, EU I, EU II and ISICUL; Fig. 3B) could indicate mixing or the presence of slumped deposits, but these were ruled out since \( ^{210}Pb \) profiles should also be flat, which was not the case (Fig. 2). Furthermore, \( C_{org} \) decay after burial was apparently almost negligible in most cases. Indeed, aerobic degradation of labile organic material near the surface of organic-rich sediments is usually so rapid that oxygen rarely penetrates ~2 mm into the sediment, therefore limiting \( C_{org} \) degradation (Kristensen et al., 2008).

Wetland \( C_{org} \) decay rates vary in relation with climate (temperature and moisture) and the quality (chemical composition) of the organic matter entering the system (Schesinger, 1997). Organic matter derived from vascular plants, such as marsh plants, mangroves and seagrasses (Moran and Hodson, 1989) is enriched in ligneous and phenolic compounds, which are refractory and require microbial breakdown to convert the more recalcitrant polymers to the more labile \( C_{org} \) available to consumers. Low oxygen concentrations in the sediment (i.e. reduced conditions) and high sediment accretion rates may also reduce \( C_{org} \) decay (Müller and Sues, 1979). In the upper core segments, where more marine conditions were identified, sediments accumulated allochthonous marine organic matter, including refractory organic matter transported from elsewhere; high concentrations of suspended particulate matter and repeated cycles of sedimentation and resuspension over diurnal and neap–spring tidal cycles promote high rates of organic matter remineralization within these zones (Abril et al., 1999).

Both cores in Sian Ka’an RB, as well as at least one of the cores in the other three study sites (ELS II in Jiquilisco Bay; EU I in Estero de Ulia Lagoon, and SA I in Salada Lagoon) showed decreasing \( C_{org} \) concentrations with increasing \(^{13}C\) values (more marine conditions; Fig. 7). The overall assessment of the data set provided a significant inverse correlation (\( p < 0.05 \)) suggesting a negative impact of marine transgression on \( C_{org} \) Concentrations. Indeed, other studies suggest that \( C_{org} \) sequestration in coastal environments is likely to decline with sea level rise (Van de Broek et al., 2016), mostly due to the reduction of terrestrial organic matter supply. However, it could be also due to a higher input of inorganic particles, such as clastic material from sediment resuspension, or biological remains such as the silica frustules of diatoms or the calcium carbonate shells of coccolithophorids. In addition, flooding may enhance organic matter decay rates even in water-logged soils (Kirwan et al., 2013) and salinity increases \( C_{org} \) microbial decomposition rates in tidal wetland sediments (Morrissey et al., 2014).

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<th>Core</th>
<th>( C_{org} ) stock (Mg ha(^{-1})) by depth and time, in sediment cores from tropical salt marshes in El Salvador and Mexico.</th>
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<td>Core</td>
<td>( C_{org} ) stock by depth (Mg ha(^{-1}))</td>
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<tr>
<td>ELS I</td>
<td>3.9 ± 0.1</td>
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<td>ELS II</td>
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<td>SA II</td>
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<tr>
<td>EU I</td>
<td>21.2 ± 0.7</td>
</tr>
<tr>
<td>EU II</td>
<td>21.2 ± 0.7</td>
</tr>
<tr>
<td>PUPA</td>
<td>6.8 ± 0.3</td>
</tr>
<tr>
<td>Min</td>
<td>3.0 ± 0.1</td>
</tr>
<tr>
<td>Max</td>
<td>65.8 ± 0.8</td>
</tr>
</tbody>
</table>

Min and max = minimum and maximum values, respectively, taking into account all the observations of all sediment cores.

a Extrapolated data up to 1 m depth, assuming negligible \( C_{org} \) degradation below 40 cm depth (see text).
b Equivalent of \( CO_2 \) = potential carbon emissions from alteration to the environment based on \( C_{org} \) stock at 1 m depth (Howard et al., 2014). Stock ratio = \( C_{org} \) stock in 50 years/\( C_{org} \) stock in 100 years.
C<sub>org</sub> burial rates (Table 1) and C<sub>org</sub> stocks (Table 2) showed very different values among and within the study sites, which highlights the complex diversity nature of tropical salt marshes. In fact, both the minimum and maximum values of C<sub>org</sub> burial rates and C<sub>org</sub> stocks (at 1 m) were found in Sian Ka’an (Table 1) and Jiquilisco Bay (Table 2), respectively. C<sub>org</sub> burial rates, highly variable in these tropical salt marshes, can reach considerably higher rates than those of peatlands from the northern hemisphere, and are comparable to the ranges reported for mangrove and salt marsh areas around the world (Table 3).

Most cores in the present study (except ISCUL, Sian Ka’an BR) showed steadily increasing C<sub>org</sub> burial rates between the 1900s and 1950s; more abrupt and intermittent changes followed, and finally increases (except SAI and SAIL cores from Salada Lagoon, Fig. 5). Since in most cores the C<sub>org</sub> concentration depth profiles did not show significant increases associated with marine transgression (Fig. 3), we consider that C<sub>org</sub> burial rates in our study sites are strongly influenced by changes in mass accumulation rates, which could be related to the impact of sea level rise, but also be affected by human interventions. For instance, as already noted (Amaya-Monterrosa et al., 2014) an abrupt decrease in MAR in the early 1950s in the cores of Jiquilisco Bay was related to the construction of dams along the Lempa River, and more recently the increase in MAR might be related to sea-level rise, but also of greater erosion in the drainage basin related to changes in land use and agricultural activities. In Salada Lagoon, the lower C<sub>org</sub> burial rates since the middle 1980s might appear to be attributable to increasing marine conditions; however, this decrease might be also influenced by the higher-energy conditions caused by the construction of the channel for the cooling waters from the nuclear power plant, which reduced la- goon inflting, and therefore MARs.

C<sub>org</sub> stock range (10–164 Mg ha<sup>−1</sup>) in the upper 30 cm) was comparable to C<sub>org</sub> stocks (30 cm depth) for mangrove (57 ± 6 Mg ha<sup>−1</sup>) and salt marsh (49 ± 5 Mg ha<sup>−1</sup>) areas in Florida, for which the authors report no significant differences (Doughty et al., 2016). In addition, most C<sub>org</sub> stocks (1 m depth, extrapolated; Table 2) were i) comparable to 355 Mg ha<sup>−1</sup> in Delaware Bay salt marshes (O’Hara et al., 2016), ii) within the range for mangrove sites in marine (271–572 Mg ha<sup>−1</sup>) and estuarine (100–315 Mg ha<sup>−1</sup>) areas in Indonesia (Weiss et al., 2016), and iii) similar to freshwater mangroves (101–491 Mg ha<sup>−1</sup>) on the Mexican coast of the Gulf of Mexico (Kaufman et al., 2016). All these results confirm the importance of mangrove and tropical salt marsh sediments as C<sub>org</sub> reservoirs, in either freshwater or coastal environments. The potential CO<sub>2</sub> emissions resulting from a given change in land use were calculated by multiplying C<sub>org</sub> stocks by 3.67, the molecular weight ratio of CO<sub>2</sub> to C (Pendleton et al., 2012); for the C<sub>org</sub> stocks determined in the present study, they varied from 110 to 1706 Mg of potential CO<sub>2</sub> emissions per hectare.

Our results show that C<sub>org</sub> storage varies considerably within a single type of environment, and also with time. In general, salt marsh carbon storage depends upon a balance between carbon inputs and decomposi-tion (Kirwan and Blum, 2011), and on local wetland conditions such as hydrology (Doughty et al., 2016). Thus, a reliable quantification of C<sub>org</sub> stocks requires a good understanding of i) the tropical salt marsh set-tings, including the energy of the environment and, consequently, the grain size of the sediments, and ii) the impacts of human interventions.

These results indicate that enhanced marine conditions caused by sea level rise promoted lower C<sub>org</sub> sediment concentrations, but increased C<sub>org</sub> burial rates, although these can also be strongly influenced by changes of sediment accretion rates caused by anthropogenic impacts. According to Figueiredo et al. (2014), the relationship between sea level rise and sediment accretion rates might be masked by local changes such as deforestation, expansion of farming land, dredging of channels, and increased paving of roadways; this is compatible with the finding by Mariotti and Fagherazzi (2013) that sediment starvation produced by river dredging and damming is a major anthropogenic driver of marsh loss and generates effects at least comparable to the ac-celerating sea-level rise due to global warming.

6. Conclusions

Table 3

<table>
<thead>
<tr>
<th>Study site</th>
<th>C&lt;sub&gt;org&lt;/sub&gt; burial rate (g m&lt;sup&gt;−2&lt;/sup&gt; year&lt;sup&gt;−1&lt;/sup&gt;)</th>
<th>Reference</th>
</tr>
</thead>
<tbody>
<tr>
<td>Mangroves</td>
<td></td>
<td></td>
</tr>
<tr>
<td>Global mean</td>
<td>100</td>
<td>twilley et al., 1992</td>
</tr>
<tr>
<td>Global mean</td>
<td>174 ± 23</td>
<td>alongi, 2014</td>
</tr>
<tr>
<td>Estimated global value</td>
<td>167</td>
<td>alongi, 2012</td>
</tr>
<tr>
<td>Estimated global value</td>
<td>139</td>
<td>duarte et al., 2005</td>
</tr>
<tr>
<td>Mean value from Gulf of Mexico, Pacific and Indian Ocean</td>
<td>188 ± 142&lt;sup&gt;a&lt;/sup&gt;</td>
<td>chmura et al., 2003</td>
</tr>
<tr>
<td>Revision of global burial rate</td>
<td>115&lt;sup&gt;b&lt;/sup&gt;</td>
<td>bouillon et al., 2008</td>
</tr>
<tr>
<td>Global geometric mean at the local scale&lt;sup&gt;c&lt;/sup&gt;</td>
<td>165 (±40, −31)</td>
<td>breithaupt et al., 2012</td>
</tr>
<tr>
<td>Riverine area (past 50 years)</td>
<td>162 ± 5</td>
<td>marchio et al., 2016</td>
</tr>
<tr>
<td>Fringe area (past 50 years)</td>
<td>127 ± 6</td>
<td>marchio et al., 2016</td>
</tr>
<tr>
<td>Everglades National Park</td>
<td>151–168</td>
<td>snoak et al., 2013</td>
</tr>
<tr>
<td>Salt marshes</td>
<td></td>
<td></td>
</tr>
<tr>
<td>Estimated global value</td>
<td>151</td>
<td>duarte et al., 2005</td>
</tr>
<tr>
<td>(geometric mean)</td>
<td></td>
<td></td>
</tr>
<tr>
<td>Northern hemisphere mean value</td>
<td>218 ± 235&lt;sup&gt;a&lt;/sup&gt;</td>
<td>chmura et al., 2003</td>
</tr>
<tr>
<td>Mean value (n = 96 sites)</td>
<td>218 ± 24</td>
<td>melcoid et al., 2011</td>
</tr>
<tr>
<td>Peatlands</td>
<td></td>
<td></td>
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<tr>
<td>Boreal areas</td>
<td>29 ± 13</td>
<td>Mitsch et al., 2013</td>
</tr>
<tr>
<td>Canada</td>
<td>20–30</td>
<td>roulet, 2000</td>
</tr>
<tr>
<td>Europe and North American boreal areas</td>
<td>23</td>
<td>dean and gorham, 1998</td>
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<tr>
<td>Finland boreal areas</td>
<td>26</td>
<td>turunen et al., 2002</td>
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<tr>
<td>Finland Arctic areas</td>
<td>17</td>
<td>turunen et al., 2002</td>
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<tr>
<td>Scottish inundated areas</td>
<td>42</td>
<td>whittle and gallego-Sala, 2016</td>
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<tr>
<td>Tropical salt marshes</td>
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<td>Jiquilisco Bay, El Salvador</td>
<td>3–255</td>
<td>this study</td>
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<td>Estero de Urais Lagoon, Mexico</td>
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<td>Salada Lagoon, Mexico</td>
<td>3–94</td>
<td>this study</td>
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<tr>
<td>Sian Ka’an Biosphere</td>
<td>3–383</td>
<td>this study</td>
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<td>Reserve, Mexico</td>
<td></td>
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</tbody>
</table>

<sup>a</sup> Calculated from the original reference.
<sup>b</sup> Confidence interval (95%).
Salt marshes are areas of high Corg accumulation, and that Corg decomposition within the sediments is small. In general, marine transgression catches the sedimentary Corg transported by tidal currents and dilution by a higher input of inorganic particles. In addition, Corg burial rates can be significantly influenced by changes in mass accumulation rates induced by increasing tidal flooding, but also by human interventions (e.g. higher MAR owing to erosion from the catchment as a result of land use changes, or lower MAR due to reten-

ded among all the cores were 3 and 378 g m\(^{-2}\).

Our results confirmed that tropical salt marshes are areas of high Corg accumulation, and that Corg decomposi-
tion is small. In general, marine transgression was associated to lower Corg sediment concentrations, likely caused by the accumulation of refractory sedimentary Corg transported by tidal currents and dilution by a higher input of inorganic particles. In addition, Corg burial rates can be significantly influenced by changes in mass accumulation rates induced by increasing tidal flooding, but also by human interventions (e.g. higher MAR owing to erosion from the catchment as a result of land use changes, or lower MAR due to reten-
ded among all the cores were 3 and 378 g m\(^{-2}\)).

We conclude that, although Corg storage in tropical salt marshes is substantial, it can be reduced by global climate change; this emphasizes the importance of protecting these environments to preserve their capacity to mitigate global warming.

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Appendix A. Supplementary data

Supplementary data to this article can be found online at https://doi.org/10.1016/j.scitotenv.2018.02.246.

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